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**Intermediate ice scour disturbance is key to maintaining a peak in biodiversity within the shallows
of the Western Antarctic Peninsula**

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Figure 1

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Abstract

Climate-related disturbance regimes are changing rapidly with profound consequences for ecosystems. Disturbance is often perceived as detrimental to biodiversity; however, the literature is divided on how they influence each other. Disturbance events in nature are diverse, occurring across numerous interacting trophic levels and multiple spatial and temporal scales, leading to divergence between empirical and theoretical studies. The shallow Antarctic seafloor has one of the largest disturbance gradients on earth, due to iceberg scouring. Scour rates are changing rapidly along the Western Antarctic Peninsula because of climate change and with further changes predicted, the Antarctic benthos will likely undergo dramatic shifts in diversity. We investigated benthic macro and megafaunal richness across 10-100 m depth range, much of which, 40-100m, has rarely been sampled. Macro and megafauna species richness peaked at 50 - 60m depth, a depth dominated by a diverse range of sessile suspension feeders, with an intermediate level of iceberg disturbance. Our results show that a broad range of disturbance values are required to detect the predicted peak in biodiversity that is consistent with the Intermediate Disturbance Hypothesis, suggesting ice scour is key to maintaining high biodiversity in Antarctica's shallows.

Introduction

Disturbance events occur in almost all natural ecosystems and tend to be a significant driver, influencing assemblage diversity, structure and function¹. However, the disturbance literature is divided. Some studies conceptualise disturbance as departures from a 'stable' state^{2,3} and detrimental to biodiversity⁴, whereas, others present cases where disturbance maintains high biodiversity⁵ and promotes resilience to further change⁶. This paradox can be addressed with the Intermediate Disturbance Hypothesis (IDH), which posits stable coexisting states under "intermediate" disturbance conditions where species diversity is predicted to be highest^{7,8}. The IDH itself however is disputed on both theoretical and empirical grounds, with studies rarely finding the predicted peaked relationship^{9,10}. Literature that has found evidence for peaks in diversity include successional, post-iceberg disturbance studies^{11,12} and theoretical models¹³. In contrast, a meta-analysis of disturbance studies found that a key factor in the detection of species richness peak was the inclusion of a broad range of disturbance levels, which has not been achieved in the majority of empirical studies^{14,15}. The Western Antarctic Peninsula has one of the largest disturbance gradients on earth¹⁶ and is considered a hotspot of benthic diversity¹⁷⁻¹⁹, making it an ideal natural laboratory for analysing the relationship between disturbance and diversity.

The shallow Antarctic seafloor (<40m depth) is home to one of the most naturally disturbed assemblages, due to frequent iceberg scouring disturbance^{16,20}. Ice scour disturbance, defined as when the keel of an iceberg impacts the seafloor, are distinct events in both time and space²¹ resulting in high mortality of >98.5% for macro and megafauna^{22,23}. Here, we consider any contact of ice with the seafloor that results in scour as ice scour disturbance, the majority of disturbance recorded here are likely caused to be small to large ice bergs²⁴. The frequency of ice scour disturbances varies due to bathymetry, latitude and topography with the highest frequency in the shallows; at some sites >35% seabed is scoured per year at 5 m depth²⁵. Typically ice scours are limited to ~500 m depth, though they may rarely occur deeper^{26,27}. Ice scour is the key factor driving biodiversity and structure in the Antarctic shallows^{21,28-31}. However, its influence has been little explored between 40 m and 100 m despite this depth range being an area of significant change in ice scour frequency³², so a broader study between 10 to 100m depth is required.

In recent decades, there have been drastic shifts in the cryosphere through atmospheric and marine warming due to greenhouse gas-driven climate change³³⁻³⁶. This is particularly true along the

Western Antarctic Peninsula (WAP)³⁷, a hotspot of regional physical change³⁸. In the Western Antarctic the seasonal sea-ice maximum area and duration have reduced over the last four decades³⁹ (although the signal is noisy). As a result, there has been an increase in iceberg movement (because of less time locked into seasonally frozen sea ice), increasing the frequency of ice scour impacts (~0.6 scours for each day of sea ice loss at 10 m depth)³². Increasing numbers of glaciers and ice shelves in retreat (87% along the WAP)³³, have led to high rates of iceberg calving³², where rates of ice scour across all depth ranges are likely to increase substantially over the next century^{32,40}. Longer-term predictions estimate there will be an eventual decrease in ice scour events as glaciers pass the grounding line and retreat onto land⁴⁰⁻⁴³.

Understanding how marine ice losses and ice scour will change the ecology of the Antarctic benthic macrofauna is key to understanding the future of this ecosystem¹, and provides insights into disturbance ecology. Disturbance is a heavily debated topic, and despite progress in this field, there is a lack of consensus on how this impacts systems when disturbance ranges move outside the historical norms⁴⁴. It is proposed through the Intermediate Disturbance Hypothesis that within a broad range of disturbance, species richness is maximised at intermediate levels due to competitively inferior, disturbance-tolerant species and competitively dominant, disturbance-sensitive species coexisting^{7,8}. However, many reports, which have been critical of the Intermediate Disturbance Hypothesis, only test the diversity-disturbance relationship across a small range of potential disturbance values^{14,15} or struggle to isolate relative, legacy and absolute disturbance^{2,25}. Therefore, sampling macro and megafaunal assemblages across one of the largest disturbance gradients on Earth, occurring over a small spatial scale, provides an ideal opportunity to test Intermediate Disturbance Hypothesis, and investigate relationships between disturbance and biodiversity. Furthermore, the fauna itself is data poor, between 40-100 m depth, probably due to poor overlap of sampling methods at this depth range³². Gathering comprehensive data from this assemblage before further climate-driven disturbance change is essential, if we are to understand the impacts of long-term change in this environment.

We surveyed benthic macro and megafaunal samples across 100 m depth from three sites on a steeply sloping marine rocky shore on Adelaide Island, WAP (67° 35' S, 068° 07'W, Supplementary materials, Figs S1). Most Antarctic species are relatively long-lived with extremely slow growth, reproduction and movement when compared to lower latitudes^{45,46}. It follows that these taxa are particularly good indicators of ice scour disturbance, with some recovery times predicted to be decades long (although exception exists⁴⁷). The broad ranges of disturbance regimes provide an opportunity to test disturbance-biodiversity relationships, within a similar environment and provide insights into the likely fate of the Antarctic benthos as they undergo dramatic disturbance changes over the next century. In this study, we aim to describe the patterns in macro and megafauna biodiversity from 10 to 100 m depth using multivariate analysis and then compare multiple diversity indices against the disturbance gradient, alongside multiple other environmental variables using multiple regression modelling. If ice scour is a driving influence behind biodiversity within the shallow Antarctic benthos, linear and polynomial regressions will be used to assess with the disturbance-biodiversity relationships are congruent with the IDH.

Methods

Study Area: The study area was steeply sloping rocky shores (67° 35' S, 068° 07'W) around Ryder Bay, Adelaide Island, Western Antarctic Peninsula between 10-100 m depths. Three sites were selected along the North coast of Ryder Bay, with similar topography (Supplementary Materials S2) and exposure to predominant current flow and iceberg scour, providing homogenous conditions. Adjacent to these sites, the Rothera Time Series (RaTS)⁴⁸ provided long-term (since 1997)

oceanographic measurements across all sample depths including light levels, temperature, salinity and standing stocks of phytoplankton.

Ice Scour: Ice scour is directly measured in the shallows around Rothera and Carlini stations in Antarctica, but the density of deeper scours is surveyed using ship-borne multibeam echo sounding. Where measured, ice scour occurrences are high^{1,6,7,9} and there has been a dramatically increased shift in density and/or frequency within the top 100 m^{21,27,29}. Our ice scour counts were collected through analysis of scours per square kilometre in multibeam bathymetry from the JR17001 (ICEBERGS1) cruise around Ryder Bay⁴⁴, between 0-500 m depth. Raw counts showed large variations in absolute values. Therefore, a log transformation was used to constrain the data range. An asymptotic regression curve (supplementary material, S3) provided the best fit for the data. Ice scour disturbance values between 10-100 m were then interpolated from this regression model.

Environmental factors: Environmental variables were collated from the Rothera time series (RaTS). As Antarctic macro and megafauna can be very long lived^{45,46}, this RaTS long-term data were used to describe the ambient environment experienced by the study taxa. All RaTS data were averaged across month to ensure even representation of the annual variation from 2011 to 2018. Maximum temperature range was calculated as the maximum and minimum recorded temperature from all 7 years at each specific depth. Benthic growth was calculated from bryozoa and serpulidae (spirobid worm) growth ring analysis⁴⁹ from 5-500 m depth. Bryozoa growth is considered to represent a median value for growth across all benthic taxa⁵⁰. A quadratic spline curve provided the best fit for the data; from this, we interpolated values for each 10 m depth interval across our study area (Supplementary material on spline regression, S4).

Macrofauna: Samples were collected at every 10 m depth interval between and including 10-100 m depth from 3 sites along Ryder Bay for a total of 30 stations. At each site the macrofauna assemblage and substrate were surveyed between February 2016 and June 2016, through 50 replicate images per station recorded via ROV, giving 1,500 samples in total. A modified DeepTrekker DTG2 was used to collect images and sample morphotypes (more details in supplementary material S5). Species accumulation curves were constructed for each station to ensure representation of rare species.

For each sample, a random area of seabed was selected and photographed (approximately 1.5 m²). Images were corrected for lens distortion with *Hugin*s photo editing software and cropped to remove areas with insufficient detail or those that were beyond the focal plane of the image. Macrofauna within the image were counted and identified into morphotypes. Specimens collected were later used to aid species identification and increase taxonomic resolution (188 specimens collected). Sample area could not be quantified as the seafloor was not uniform in shape, structure or composition. Attempts were made to ensure sampling was as uniform as possible and all images were scaled using two lasers but there remains an unquantifiable variability across each sample.

Data Analysis: Biodiversity was expressed as species richness, the number of macrofaunal species present within a sample, Shannon-Weiner index⁵¹ and Fisher's α ⁵². Shannon-Weiner and Fisher's α were analysed as Shannon-Weiner includes an evenness measure and Fisher's α is independent of sample size, to ensure that neither evenness nor sample size significantly alter the results. We performed linear and polynomial (quadratic and cubic) regression analyses to determine the best-fit shape of biodiversity-disturbance relationship. Variance Inflation Factors (VIF) were used to identify any collinearity (VIF values between 1-5 = moderately correlated and >5 = highly correlated⁵³). Parameters of regression were estimated using R package *lme4* with Loess smoothing using the R

package *ggplot2* to assess potential nonlinearity between biodiversity and disturbance. All statistical analyses were performed using R 3.5.2 and Minitab 19.

Macrofauna composition was analysed using Primer 7 (version 7.0.17). Taxa abundance was transformed using square root function to reduce the influence of hyper-abundance and non-metric multidimensional scaling (nMDS), using a Bray-Curtis resemblance matrix was used to compare macrofaunal composition across all depths and sites. SIMPER (SIMilarity PERcentages) analysis was used to calculate the contribution of each taxa to group similarity, across the different factor levels.

Results

Depths between 10-30 m were dominated by mobile grazers such as by *Nacella concinna* (limpets) and *Sterechinus neumayeri* (sea urchins). These depths were also coincident with the highest prevalence of algae, although coralline algae was still found in reasonably high frequencies at 60 m depth. Between 40-50 m depth, a mixed assemblage of sessile suspension feeders and mobile grazers/scavengers were dominant with species such as *Cnemidocarpa verrucosa* (solitary ascidian), *S. neumayeri* and *Ophionotus victoriae* (brittle star). At 60-100 m depth, sessile suspension feeders dominated with some associated fauna, groups of Porifera and Bryozoa in particular. Due to bryozoans only being identifiable to species level under a microscope, multiple collections were made and found two bryozoan morphotypes represent multiple species. Bryozoan diversity is likely under reported but did coincide with the species richness peak between 50-60 m depth. Suspension feeders included, *Neofungella* sp. (Stenoleamate byrozoan), *Perkinsiana littoralis* (feather worms) and *Anoxyclalyx joubeni* (structure-forming hexactinellid sponge).

No clear zonation was observed between 10-100 m depth; rather a gradual shift between assemblages with a broad overlap in species ranges (Supplementary materials Figs S6). Across all depths the assemblage composition showed large degree of variability or 'patchiness', typical of Antarctic benthos and the resulting from spatial heterogeneity in iceberg scours⁵⁴. Gastropoda, Asteroidea and Anthozoa groups showed no depth trend with individual taxa having wide depth ranges, although Actiniaria (sea anemones) tended to be found deeper (>60 m depth, but heavily species specific). Bryozoa, Ascidia and Porifera were found deeper, with the exception of *Beania* sp. (Ctenostomata Bryozoa) and *Cnemidocarpa verrucosa*. *Sterechinus neumayeri* and *O. victoriae* had a notable prevalence across all depths, although these taxa were found in higher abundance at depths shallower and deeper than 50 m, respectively. Representatives of the Holothuroidea (sea cucumbers), Hydrozoa and Entoprocta were more prevalent at intermediate depths (30-70 m).

Counts of scours per square kilometre on seabed mapping (vessel multibeam) data spanning Marguerite Bay showed that ice scour disturbance varied considerably across all depths. Scour density decreased from 1.75×10^5 scours per square kilometre at 10 m depth to 3.92 scours per square kilometre at 100 m depth. Species richness showed a peaked relationship with study depth. We found an average species richness of 5.77 per image at 10 m depth, increasing to 22.49 between 50-60 m depth, before decreasing to 14.77 species richness by 100 m depth. The peak in species richness coincided with 32 scours per square kilometre. Linear and polynomial regression analysis found a cubic function ($F_{3,1496} = 385.94$, $r^2 = 0.44$, $p < 0.01$) and provided the best-fit relationship between biodiversity and disturbance (Figure 1.). The regression line shows a clear unimodal relationship, with a wide range of species richness at each level of disturbance. The maximum range of species richness at each depth was on average, 28.2 species (average Standard Deviation 5.43, across all depths). We found similar diversity-disturbance trends with all diversity indices (Supplementary materials Figs S7). As Shannon-Weiner ($r^2=32.6$, $p < 0.001$) and Fisher's α ($r^2=26.7$,

213 $p < 0.001$) diversity indices had lower r^2 values than the relationship between depth and species
 214 richness ($r^2 = 48.7$, $p < 0.001$), further analyses used species richness.

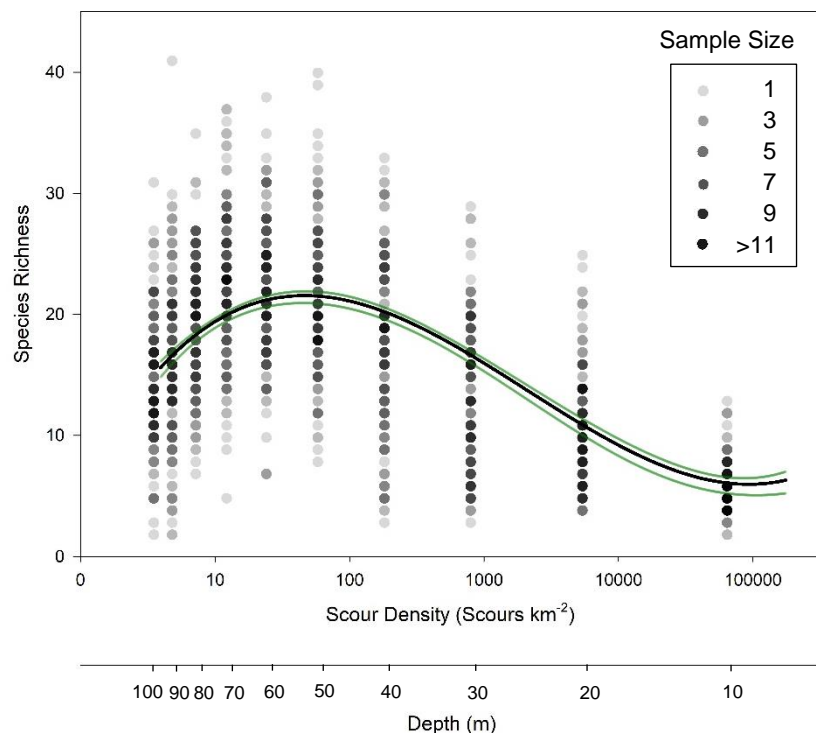


Figure 1. Relationship between species richness and disturbance.

Line (model) of best fit was non-linear regression (cubic model, black line). Points are samples, with increasing shades of grey representing a greater number of samples. Green lines are 95% confidence intervals. Total sample number is 1500, evenly divided across 10 m depth intervals. Plot constructed in RStudio v1.1.463, <https://www.rstudio.com/>.

215 Growth data for selected macrofauna interpolated from the literature⁴⁹ rose from $0.08 \text{ g day}^{-1} \text{ m}^{-2}$ at
 216 10 m to peak at $0.14 \text{ C g day}^{-1} \text{ m}^{-2}$ at 40 m depth, decreasing to $0.12 \text{ g day}^{-1} \text{ m}^{-2}$ at 100 m depth
 217 (supplementary material S8). Growth correlates with ice scour disturbance ($VIF = 1.68$) and therefore
 218 has a quadratic correlation with species richness ($F_{2,1497} = 291.20$, $r^2 = 0.279$, $p < 0.001$). However, the
 219 maximum range of values for growth between 40-100 m depth, was $0.02 \text{ g day}^{-1} \text{ m}^{-2}$. These values
 220 are below the signal noise threshold, of $0.05 \text{ g day}^{-1} \text{ m}^{-2}$, and cannot be distinguish from experimental
 221 error. Average annual salinity varied by a maximum 0.54 ‰ across all stations, which is in line with
 222 previous work on coastal Southern Ocean salinity being stable and constant throughout the year
 223 (except in the intertidal zone)⁵⁵ (supplementary material S5). The range of growth and salinity were
 224 not considered large enough to detect any correlation with species richness, so were removed from
 225 the analysis.

226 Average annual sea temperature was -1.04°C at 10 m depth. This variable decreased to a minimum
 227 of -1.09°C at 25 m depth, before increasing to -0.73°C at 100 m (supplementary material S3).
 228 Average annual sea temperature was correlated with ice scour disturbance ($VIF = 1.78$) but did not
 229 correlate with species richness. Maximum sea temperature range at 10 m depth was 4.00°C , which
 230 decreased exponentially with depth, reaching 2.71°C at 100 m depth (supplementary material S3).
 231 Chlorophyll α concentration decreased at an exponential rate with depth from 1.85 mg m^{-3} at 10 m
 232 to 0.16 mg m^{-3} at 100 m, as did photosynthetically active radiation, from $47.70 \text{ } \mu\text{mol m}^{-2} \text{ s}^{-1}$ to 0.18
 233 $\mu\text{mol m}^{-2} \text{ s}^{-1}$ (supplementary material S3). Sea temperature range, chlorophyll α and light levels
 234 exponentially decreased with depth, with the majority of change occurring in the top 20 m depth. All
 235 variables had a strong collinearity with scour density ($VIF = 20.09, 205.47, 25.24$ respectively) and
 236 were therefore removed from the model. Linear and polynomial regression analyses for sea
 237 temperature range, chlorophyll α and light levels had a similar unimodal relationship, as ice scour
 238 disturbance. However, all environmental variables had lower r^2 values and poorer overall fit,
 239 particularly past 30-40 m depth. In addition, there were only small differences between sites, and

the inclusion of sediment and site did not significantly improve the model (Supplementary information on multiple regression analysis S4). These were tested to account for variation in ice abundance, topography and current between all 3 sites.

Discussion

The Antarctic marine shallows are home to one of the largest natural disturbance gradients on earth, up to 100% mortality across the entire macrobenthic population within the intertidal (with some exceptions see Waller, et al. ⁵⁶), to near 0% mortality²¹ around 200 m depth⁵⁷. Shallower than 40 m depth ice scour disturbance is a key controlling factor^{21,28-30} as only disturbance resilient species are able to persist, reducing species richness^{7,8}. However between 40-100m depth there is little information on which environmental factors influence the Antarctic benthos and furthermore what species occupy this depth range³². Deeper than 40 m we found a unimodal relationship between macro and megafauna species richness and ice scour disturbance, with a peak in species richness at intermediate levels of ice scour disturbance. This concurs with the Intermediate Disturbance Hypothesis, a widely recognised concept, but one that has produced many reviews and critiques¹⁵.

The disturbance-diversity pattern identified across our depth range showed an extreme variability in species richness across all depths. This patchiness is suggestive of ice scour disturbance being the driving factor, as a spatially and temporally discrete mass mortality event^{22,23}. The variation in species richness amongst samples from similar depth likely reflect a patchwork of assemblages at different stages of recovery, from previous ice scour events. However, 'patchiness' (or spatial heterogeneity) was lowest at 10 m depth, which was dominated by a mobile assemblage, which could rapidly re-invade recent iceberg scours, the impact of ice scour impacts across a wider area, enough to homogenise the fauna at this depth²⁵.

The influence of other environmental variables could not be completely isolated from disturbance, although many of them showed minor changes beyond 30 m depth. Additionally, we do not know at what depths lower thresholds of disturbance are reached and species richness starts to be controlled by other factors. Likely the flux of food particles from the surface, which much of the Antarctic seafloor community is reliant on⁵⁸, will become a crucial factor at depth. For example, Jansen, et al. ⁵⁹ showed that the abundance and richness of types of benthic fauna could be predicted by food availability at depths below 200 m. We could not confirm any influence of light level or chlorophyll a concentration on biodiversity; however, they are likely to play a major, but perhaps complex, role in the structuring of benthic biota and ecosystem dynamics⁶⁰, particularly below the depth of peak biodiversity.

The Western Antarctic Peninsula is a climate change hotspot that is predicted to warm if current emissions continue⁶¹. This change is also likely to result in a profound impact on ice scour disturbance, as glaciers continue to retreat and sea-ice reduces in both extent and duration^{33-35,41}. Over the next century icebergs are likely to calve at an increased rate and with higher mobility as they are less likely to be held in place by seasonal sea ice⁴⁰. As argued in this study, ice scour disturbance is a key controlling factor down to 100 m depth; if disturbance regimes continue to change, we expect benthic biodiversity to alter considerably.

We suggest two potential futures within the next century for biodiversity in the shallows, based on the diversity-disturbance patterns reported in this study and the current composition of the Antarctic macro and mega-fauna. First, if scour disturbance increases rapidly the macro and megafaunal assemblage will struggle to redistribute, particularly if these species ranges are restricted by depth-dependent environmental and biological factors. The majority of macro and megafaunal species are long lived with slow growth, locomotion and reproduction, when compared

to lower latitudes^{45,46} (but may grow faster with moderate warming⁶²). Within this context, a century may not be long enough for these species to migrate away from, or adapt to, new conditions. Increasing ice scour is expected to remove many of the competitively-dominant, disturbance-sensitive species, such as *Mycale acerata* (sponge), which have slow growth and reproduction rates⁶³. However, many macro and megafauna species have wide depth-ranges (*M. acerata* for example between 20-90 m depth) and so although species richness is controlled by ice scour in the shallows, species may still exist at extremely low frequencies across a wide spectrum of disturbance levels.

The presence of species across a wide depth gradient, may allow a few individuals found at the extremes of their ranges to thrive as conditions shift in their favour. The broad depth-ranges of many species support a second prediction, that the increase in ice scour disturbance would redistribute species into deeper waters, as the diversity migrates in response to a new disturbance pattern. The second prediction is based on biodiversity being driven by disturbance, even at the deeper end of our depth range. Beyond 100 m depth the relative difference in disturbance is minute and it is likely that primary production (more specifically bloom duration⁴¹) will be the limiting factor, restricting the depth over which these species can redistribute. However, the pattern between sea-surface chlorophyll and species richness is multifaceted, with trophic dependent relationships and dependent on multiple physical variables⁵⁹. This study cannot disentangle where, or if, the relative contribution of disturbance is surpassed by primary production as a driving factor and instead asserts that between 60-100 m depth the influence of disturbance is likely to wane.

With both predictions, we can expect species richness loss in the shallows (10-30 m) as disturbance tolerant species reach their limit and either redistribute to deeper waters or are extirpated. Both of the predictions made here are by no means mutually exclusive, there may well be a drop in diversity across all depth ranges, as species are unable to move outside of their established ranges, combined with a shift in the now reduced biodiversity peak, as the intermediate levels of disturbance shift deeper. The eventual limit of the depth shift in biodiversity will likely be dictated by the depth related reduction in primary production⁶⁴. However, climate change-induced sea ice changes and associated changes in light regime⁶⁵ are predicted to increase bloom duration⁴¹ potentially allowing more species to persist at a greater depth.

In particular, species such as *Sterechinus neumayeri* and *Ophionotus victoriae* both found in high abundance across a large depth range with catholic diets^{66,67}, will likely thrive as niches shift and new opportunities become available. A key feature in assemblage response to disturbance shift is dispersal capability²⁵; broadcast-spawning species, such as *Cnemidocarpa verrucosa*⁶⁸, may be better able to redistribute in response to the changing environment. While species that have low reproductive rates but are sensitive to climate forcing, such as *Anoxycalx joubini* (structure-forming hexactinellid sponge) may spawn to respond to these changing conditions⁶⁹. Generally species with low reproductive rates are likely to suffer, however this may be countered by mobile species, whose adults can adjust depth ranges through movement such as *Trematomus bernacchii* (Nototheniidae fish)⁷⁰. Ultimately however if warming continues glaciers will retreat past grounding lines and iceberg calving rates will drop dramatically resulting in a complete reversal to low levels of iceberg disturbance across all depths⁴². This will likely form a new climax community with lower diversity and dominated by porifera (sponges) usually found in deeper water, as can be seen in small, sheltered areas of the seabed where much deeper species dominate (e.g., overhangs and caves²⁸). However, in the previously high disturbance area between 10-30 m there may be small increases in richness and diversity, as macro-algae and their associated fauna increase.

To summarise, even though the Intermediate Disturbance Hypothesis is debated^{9,14,15,71}, our results are congruent with this explanation for the Antarctic benthos disturbance-diversity pattern which can be detected because of the broad range of disturbance regimes included in this study. The consequences of the diversity-disturbance patterns within shallow Antarctic benthos will have profound impacts, particularly with glacial retreat opening new fjordic habitats and potential providing new carbon sinks and negative climate feedback loops⁴². The future of the shallow Antarctic benthos is likely to involve dramatic fluctuations in biodiversity and ecosystem functioning, and should warming continue, could ultimately lead to locally large losses in biodiversity with far-reaching implications.

- 1 Dell, J. *et al.* Interaction diversity maintains resiliency in a frequently disturbed ecosystem. *Frontiers in Ecology and Evolution* **7**, 145 (2019).
- 2 White, P. S. & Pickett, S. T. A. in *The Ecology of Natural Disturbance and Patch Dynamics* (eds S. T. A. Pickett & P. S. White) 3–13 (NY: Academic Press, 1985).
- 3 Newman, E. A. Disturbance Ecology in the Anthropocene. *Frontiers in Ecology and Evolution* **7**, doi:10.3389/fevo.2019.00147 (2019).
- 4 Barnosky, A. D. *et al.* Approaching a state shift in Earth's biosphere. *Nature* **486**, 52-58 (2012).
- 5 Yuan, Z., Jiao, F., Li, Y. & Kallenbach, R. L. Anthropogenic disturbances are key to maintaining the biodiversity of grasslands. *Scientific Reports* **6**, 22132 (2016).
- 6 Hughes, A. R., Byrnes, J. E., Kimbro, D. L. & Stachowicz, J. J. Reciprocal relationships and potential feedbacks between biodiversity and disturbance. *Ecol Lett* **10**, 849-864, doi:10.1111/j.1461-0248.2007.01075.x (2007).
- 7 Connell, J. H. & Slatyer, R. O. Mechanisms of succession in natural communities and their role in community stability and organization. *The American Naturalist* **111**, 1119-1144 (1977).
- 8 Connell, J. H. Diversity in tropical rain forests and coral reefs. *Science* **199**, 1302-1310 (1978).
- 9 Fox, J. W. The intermediate disturbance hypothesis should be abandoned. *Trends Ecol Evol* **28**, 86-92, doi:10.1016/j.tree.2012.08.014 (2013).
- 10 Sheil, D. & Burslem, D. F. Disturbing hypotheses in tropical forests. *Trends in Ecology & Evolution* **18**, 18-26 (2003).
- 11 Teixidó, N., Garrabou, J., Gutt, J. & Arntz, W. E. Recovery in Antarctic benthos after iceberg disturbance: trends in benthic composition, abundance and growth forms. *Marine Ecology Progress Series* **278**, 1-16, doi:10.3354/meps278001 (2004).
- 12 Teixidó, N., Garrabou, J., Gutt, J. & Arntz, W. Iceberg disturbance and successional spatial patterns: the case of the shelf Antarctic benthic communities. *Ecosystems* **10**, 143-158 (2007).
- 13 Johst, K., Gutt, J., Wissel, C. & Grimm, V. Diversity and disturbances in the Antarctic megabenthos: feasible versus theoretical disturbance ranges. *Ecosystems* **9**, 1145-1155 (2006).
- 14 Mackey, R. L. & Currie, D. J. The Diversity-Disturbance Relationship: Is it generally strong and peaked? *Ecology* **82**, 3479-3492, doi:doi:10.1890/0012-9658(2001) (2001).
- 15 Huston, M. A. Disturbance, productivity, and species diversity: empiricism vs. logic in ecological theory. *Ecology* **95**, 2382-2396, doi:doi:10.1890/13-1397.1 (2014).
- 16 Smale, D. A., Brown, K. M., Barnes, D. K., Fraser, K. P. & Clarke, A. Ice scour disturbance in Antarctic waters. *Science* **321**, 371, doi:10.1126/science.1158647 (2008).
- 17 Griffiths, H. J., Danis, B. & Clarke, A. Quantifying Antarctic marine biodiversity: The SCAR-MarBIN data portal. *Deep Sea Research Part II: Topical Studies in Oceanography* **58**, 18-29, doi:10.1016/j.dsr2.2010.10.008 (2011).

379 18 Grange, L. J. & Smith, C. R. Megafaunal communities in rapidly warming fjords along the
380 West Antarctic Peninsula: hotspots of abundance and beta diversity. *PloS one* **8**, e77917
381 (2013).

382 19 Gutt, J., Griffiths, H. J. & Jones, C. D. Circumpolar overview and spatial heterogeneity of
383 Antarctic macrobenthic communities. *Marine Biodiversity* **43**, 481-487, doi:10.1007/s12526-
384 013-0152-9 (2013).

385 20 Potthoff, M., Johst, K. & Gutt, J. How to survive as a pioneer species in the Antarctic benthos:
386 minimum dispersal distance as a function of lifetime and disturbance. *Polar Biology* **29**, 543-
387 551 (2006).

388 21 Convey, P. *et al.* The spatial structure of Antarctic biodiversity. *Ecological Monographs* **84**,
389 203-244 (2014).

390 22 Peck, L. S., Brockington, S., Vanhove, S. & Beghyn, M. Community recovery following
391 catastrophic iceberg impacts in a soft-sediment shallow-water site at Signy Island,
392 Antarctica. *Marine Ecology Progress Series* **186**, 1-8 (1999).

393 23 Lee, H., Vanhove, S., Peck, L. & Vincx, M. Recolonisation of meiofauna after catastrophic
394 iceberg scouring in shallow Antarctic sediments. *Polar Biology* **24**, 918-925,
395 doi:10.1007/s003000100300 (2001).

396 24 Armstrong, T. World Meteorological Organization. WMO sea-ice nomenclature.
397 Terminology, codes and illustrated glossary. Edition 1970. Geneva, Secretariat of the World
398 Meteorological Organization, 1970.[ix], 147 p.[including 175 photos]+ corrigenda
399 slip.(WMO/OMM/BMO, No. 259, TP. 145.). *Journal of Glaciology* **11**, 148-149 (1972).

400 25 Robinson, B. J., Barnes, D. K. & Morley, S. A. Disturbance, dispersal and marine assemblage
401 structure: A case study from the nearshore Southern Ocean. *Marine Environmental*
402 *Research*, 105025 (2020).

403 26 Gutt, J., Starmans, A. & Dieckmann, G. Impact of iceberg scouring on polar benthic habitats.
404 *Marine Ecology Progress Series* **137**, 311-316 (1996).

405 27 Barnes, D. K. A. & Conlan, K. E. Disturbance, colonization and development of Antarctic
406 benthic communities. *Philos Trans R Soc Lond B Biol Sci* **362**, 11-38,
407 doi:10.1098/rstb.2006.1951 (2007).

408 28 Smale, D. A. Ecological traits of benthic assemblages in shallow Antarctic waters: does ice
409 scour disturbance select for small, mobile, secondary consumers with high dispersal
410 potential? *Polar Biology* **31**, 1225-1231, doi:10.1007/s00300-008-0461-9 (2008).

411 29 Barnes, D. K. A. The influence of ice on polar nearshore benthos. *Journal of the Marine*
412 *Biological Association of the United Kingdom* **79**, 401-407 (1999).

413 30 Gutt, J. On the direct impact of ice on marine benthic communities, a review. *Polar Biology*
414 **24**, 553-564 (2001).

415 31 Barnes, D. K. A. & Tarling, G. A. Polar oceans in a changing climate. *Curr Biol* **27**, R454-R460,
416 doi:10.1016/j.cub.2017.01.045 (2017).

417 32 Barnes, D. K. A., Fleming, A., Sands, C., J., Quartino, M., Liliana & Deregibus, D. Icebergs, sea
418 ice, blue carbon and Antarctic climate feedbacks. *Philosophical Transactions of the Royal*
419 *Society A: Mathematical, Physical and Engineering Sciences* **376**, 20170176,
420 doi:10.1098/rsta.2017.0176 (2018).

421 33 Cook, A. J., Fox, A. J., Vaughan, D. G. & Ferrigno, J. G. Retreating glacier fronts on the
422 Antarctic Peninsula over the past half-century. *Science* **308**, 541-544,
423 doi:10.1126/science.1104235 (2015).

424 34 Cook, A. *et al.* Ocean forcing of glacier retreat in the western Antarctic Peninsula. *Science*
425 **353**, 283-286 (2016).

426 35 Clarke, A. *et al.* Climate change and the marine ecosystem of the western Antarctic
427 Peninsula. *Philos Trans R Soc Lond B Biol Sci* **362**, 149-166, doi:10.1098/rstb.2006.1958
428 (2007).

429 36 Turner, J. & Comiso, J. Solve Antarctica's sea-ice puzzle. *Nature News* **547**, 275 (2017).

430 37 Meredith, M. P. & King, J. C. Rapid climate change in the ocean west of the Antarctic
431 Peninsula during the second half of the 20th century. *Geophysical Research Letters* **32**,
432 doi:doi:10.1029/2005GL024042 (2005).

433 38 Barnes, D. K. A. & Souster, T. Reduced survival of Antarctic benthos linked to climate-
434 induced iceberg scouring. *Nature Climate Change* **1**, 365-368, doi:10.1038/nclimate1232
435 (2011).

436 39 Parkinson, C. L. Global Sea Ice Coverage from Satellite Data: Annual Cycle and 35-Yr Trends.
437 *Journal of Climate* **27**, 9377-9382, doi:10.1175/jcli-d-14-00605.1 (2014).

438 40 Rogers, A. *et al.* Antarctic futures: An assessment of climate-driven changes in ecosystem
439 structure, function, and service provisioning in the Southern Ocean. *Annual review of marine*
440 *science* **12**, 87-120 (2020).

441 41 Morley, S. A. *et al.* Global drivers on Southern Ocean ecosystems: changing physical
442 environments and anthropogenic pressures in an Earth system. *Frontiers in Marine Science*
443 **7**, 1097 (2020).

444 42 Barnes, D. K. *et al.* Blue carbon gains from glacial retreat along Antarctic fjords: What should
445 we expect? *Global change biology* **26**, 2750-2755 (2020).

446 43 Barnes, D. K. A. in *Carbon Capture, Utilization and Sequestration* Ch. Chapter 3, (2018).

447 44 Bowler, D. *et al.* The geography of the Anthropocene differs between the land and the sea.
448 *bioRxiv*, 432880, doi:10.1101/432880 (2019).

449 45 Arntz, W., Brey, T. & Gallardo, V. Antarctic zoobenthos. *Oceanography and marine biology*
450 **32**, 241-304 (1994).

451 46 Clarke, A. Marine benthic populations in Antarctica: patterns and processes. *Antarctic*
452 *Research Series* **70**, 373-388 (1996).

453 47 Fillinger, L., Janussen, D., Lundälv, T. & Richter, C. Rapid glass sponge expansion after
454 climate-induced Antarctic ice shelf collapse. *Current Biology* **23**, 1330-1334 (2013).

455 48 Clarke, A., Meredith, M. P., Wallace, M. I., Brandon, M. A. & Thomas, D. N. Seasonal and
456 interannual variability in temperature, chlorophyll and macronutrients in northern
457 Marguerite Bay, Antarctica. *Deep Sea Research Part II: Topical Studies in Oceanography* **55**,
458 1988-2006, doi:https://doi.org/10.1016/j.dsr2.2008.04.035 (2008).

459 49 Barnes, D. K. A. Iceberg killing fields limit huge potential for benthic blue carbon in Antarctic
460 shallows. *Glob Chang Biol* **23**, 2649-2659, doi:10.1111/gcb.13523 (2017).

461 50 Pinkerton, M., Bradford-Grieve, J., Bowden, D. & Cummings, V. Benthos: trophic modelling
462 of the Ross Sea. *Supporting document to CCAMLR science* **17**, 1-31 (2010).

463 51 Pielou, E. Shannon's formula as a measurement of species diversity: it's use and disuse. *Am*
464 *Nat* **100**, 463-465 (1966).

465 52 Fisher, R. A., Corbet, A. S. & Williams, C. B. The relation between the number of species and
466 the number of individuals in a random sample of an animal population. *The Journal of*
467 *Animal Ecology*, 42-58 (1943).

468 53 Everitt, B. & Skrondal, A. *The Cambridge dictionary of statistics*. Vol. 106 (Cambridge
469 University Press Cambridge, 2002).

470 54 Smale, D. A., Barnes, D. K. A. & Fraser, K. P. P. The influence of ice scour on benthic
471 communities at three contrasting sites at Adelaide Island, Antarctica. *Austral Ecology* **32**,
472 878-888, doi:10.1111/j.1442-9993.2007.01776.x (2007).

473 55 Peck, L. S., Convey, P. & Barnes, D. K. A. Environmental constraints on life histories in
474 Antarctic ecosystems: tempos, timings and predictability. *Biological Reviews* **81**, 75-109,
475 doi:10.1017/s1464793105006871 (2006).

476 56 Waller, C., Worland, M., Convey, P. & Barnes, D. Ecophysiological strategies of Antarctic
477 intertidal invertebrates faced with freezing stress. *Polar Biology* **29**, 1077-1083 (2006).

478 57 Barnes, D. K. A. Polar zoobenthos blue carbon storage increases with sea ice losses, because
479 across-shelf growth gains from longer algal blooms outweigh ice scour mortality in the
480 shallows. *Glob Chang Biol* **23**, 5083-5091, doi:10.1111/gcb.13772 (2017).

- 58 Smith, C. R., Mincks, S. & DeMaster, D. J. A synthesis of benthic-pelagic coupling on the Antarctic shelf: food banks, ecosystem inertia and global climate change. *Deep Sea Research Part II: Topical Studies in Oceanography* **53**, 875-894 (2006).
- 59 Jansen, J. *et al.* Abundance and richness of key Antarctic seafloor fauna correlates with modelled food availability. *Nature Ecology & Evolution* **2**, 71-80 (2018).
- 60 Henley, S. F. *et al.* Changing biogeochemistry of the Southern Ocean and its ecosystem implications. *Frontiers in Marine Science* **7**, 581 (2020).
- 61 Marshall, G. J. *et al.* Causes of exceptional atmospheric circulation changes in the Southern Hemisphere. *Geophysical Research Letters* **31** (2004).
- 62 Ashton, G. V., Morley, S. A., Barnes, D. K., Clark, M. S. & Peck, L. S. Warming by 1 C drives species and assemblage level responses in Antarctica's marine shallows. *Current Biology* **27**, 2698-2705. e2693 (2017).
- 63 Riesgo, A. *et al.* Some like it fat: Comparative ultrastructure of the embryo in two demosponges of the genus *Mycale* (order poecilosclerida) from Antarctica and the Caribbean. *PloS one* **10**, e0118805 (2015).
- 64 Toszogyova, A. & Storch, D. Global diversity patterns are modulated by temporal fluctuations in primary productivity. *Global Ecology and Biogeography* **28**, 1827-1838 (2019).
- 65 Clark, G. F. *et al.* Light-driven tipping points in polar ecosystems. *Global Change Biology* **19**, 3749-3761 (2013).
- 66 Brockington, S., Clarke, A. & Chapman, A. Seasonality of feeding and nutritional status during the austral winter in the Antarctic sea urchin *Sterechinus neumayeri*. *Marine Biology* **139**, 127-138 (2001).
- 67 Fratt, D. B. & Dearborn, J. Feeding biology of the Antarctic brittle star *Ophionotus victoriae* (Echinodermata: Ophiuroidea). *Polar Biology* **3**, 127-139 (1984).
- 68 Sahade, R., Tatián, M. & Esnal, G. B. Reproductive ecology of the ascidian *Cnemidocarpa verrucosa* at Potter Cove, South Shetland Islands, Antarctica. *Marine Ecology Progress Series* **272**, 131-140 (2004).
- 69 Dayton, P. K. *et al.* Recruitment, growth and mortality of an Antarctic hexactinellid sponge, *Anoxycalyx joubini*. *PloS one* **8**, e56939 (2013).
- 70 Vacchi, M., Cattaneo-Vietti, R., Chiantore, M. & Dalù, M. Predator-prey relationship between the nototheniid fish *Trematomus bernacchii* and the Antarctic scallop *Adamussium colbecki* at Terra Nova Bay (Ross Sea). *Antarctic Science* **12**, 64-68 (2000).
- 71 Sheil, D. & Burslem, D. F. Defining and defending Connell's intermediate disturbance hypothesis: a response to Fox. *Trends Ecol Evol* **28**, 571-572, doi:10.1016/j.tree.2013.07.006 (2013).

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Figure legends:

527 **Figure 1. Relationship between species richness and disturbance.** Line (model) of best fit was non-
528 linear regression (cubic model, black line). Points are samples, with increasing shades of grey
529 representing a greater number of samples. Green lines are 95% confidence intervals. Total sample
530 number is 1500, evenly divided across 10 m depth intervals. Plot constructed in RStudio v1.1.463,
531 <https://www.rstudio.com/>.